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# Forest Owners' Acceptance of Incentive Based Policy Instruments in Forest Biodiversity Conservation – A Choice Experiment Based Approach

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Finland has launched a new policy programme (METSO) to enhance conservation of forest biodiversity. In non-industrial private forests, the policy is based on economic incentives and voluntarism on the part of forest owners. While biodiversity conservation is the main target of the policy, social acceptability is considered to be of great importance. This study examined the factors that affect the acceptability of biodiversity conservation contracts among private forest owners, and the amount of compensation needed to keep the forest owners at least as well off as before the contract. Choice experiment method was used to analyse the data that were collected by surveying 3000 Finnish private forest owners. Analysing separately those respondents who were willing to enter into a conservation contract allowed an assessment of the impact of forest owners' heterogeneity on compensation amount. The results show how the welfare of forest owners shifts when the contract terms are changed. In a base scenario the forest owner was assumed to be the initiator of the contract that would require only small patches of forest to be protected, and would also bind new forest owners over its duration of ten years. For all respondents, the average demand for compensation would be around 224 euros annually. When those always choosing the "no additional conservation" alternative were excluded, the average welfare impact of the base scenario was positive.

Keywords forest owners, heterogeneity, incentive based policy mechanism, choice experiment method

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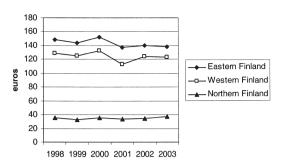
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## **1** Introduction

Forests in Finland produce a multitude of environmental services alongside consumable goods like timber and berries. Some of these goods and especially the services are public goods. Provision of public goods on private lands is not necessarily at a socially optimal level, as private decision makers might not internalise them into their objective function. The conservation of native species or biodiversity provides typical public goods, the benefit of which cannot be exclusive to the private forest owner.

The definition of the ownership of different forest goods and services is specific to national jurisdiction, and thus varies between countries. Depending on the definition of ownership, environmental public goods can be viewed as either positive or negative externalities. If the property rights were complete and exclusive covering all the aspects of forestland, any conservation values provided in the forest would be positive externalities. Alternatively, a national law might forbid a landowner from reducing conservation values, thereby implying that the property rights for the service reside with society. According to Coasian logic the compensation obligation depends on the property rights. Were the land ownership complete and exclusive, the owner should be compensated for all the lost private values, e.g. timber revenue, when the resource is used to produce public services (Innes et al. 1998). In Finland, most citizens are concerned with justice toward forest owners and their sovereignty. A clear majority of citizens is in favour of full compensation to the forest owners for lost revenues and possible costs of nature conservation action, and support forest owners' sovereignty in forest management decisions (Horne 2002).

In Finland about three quarters of the land area is forested. Over a third of the threatened species are forest organisms. Management over centuries has changed the forest composition in such a way that many species that are dependent on decayed wood or old-growth forests are now threatened (Rassi et al. 2001). Currently, 7.5 per cent of productive forestland with annual yield of more than 1 m<sup>3</sup>/hectare is protected in Finland. Most of the protected areas are located in the northern part of the country where 17 per cent



**Fig. 1.** Net earnings in non-industrial private forestry per hectare in the three regions. Source: Finnish Statistical Yearbook of Forestry (Finnish... 2004).

of the forest area is protected by law. However, most of the habitats of endangered species are located in Southern Finland where only 1.8 per cent of forestland is protected. (Virkkala et al. 2000, Ruuhijärvi et al. 2000).

In order to achieve conservation goals in the long run, the social sustainability of the forest conservation policy should be assessed. One aspect of social sustainability is the general acceptance of the goals, impacts and implementation of the policy. The non-industrial private forest (NIPF) owners own 61 per cent of forests in Finland, and almost 75 per cent in the southern part of the country. State ownership in Southern Finland is less than 10 per cent. Nearly 20 per cent of Finnish households own a forest holding (Finnish... 2004). Hence, the economic and social implications of forest protection fall predominantly on this sector of society

Such forest management methods that would maximize timber revenue and biodiversity value at the same plot simultaneously do not exist, at least for boreal environment. The highest potential revenue from forestry is obtained by using intensive, economical and large scale methods with shorter rotation period (for theoretical discussion see Bowes and Krutilla 1989, for empirical results see Hyytiäinen and Tahvonen 2003). Forest revenue is an important source of income to many forest owners (Kangas and Niemeläinen 1996). About 80 per cent of forest owners place some economic expectations on their forest property (Karppinen et al. 2002). However, not all forest owners would face the same opportunity cost for biodiversity conservation. The average net forestry earnings per hectare vary according to the vegetational zones. In this study, Finland is divided into three regions with different net earnings per hectare (Fig. 1).

Also the social implications of setting forest aside for conservation purposes would depend on the individual characteristics and situation of each forest owner. Forest owners' attitudes and values towards biodiversity conservation and forest owners' sovereignty differ considerably (Horne et al. 2004). Also, about a third of forest owners harvest themselves and two thirds engage in some forestry activities, like planting, or thining young stands, while some owners have no personal involvement in management (Karppinen et al. 2002).

Conventionally, Finnish nature conservation policy has been implemented through the state buying areas with biodiversity or conservation value. The NIPF owners have not always approved these top-down approaches to nature conservation. Along with the recent trend in international biodiversity governance, there has been a shift toward incentive based policy mechanisms. In 2002, the Finnish government accepted a programme for action (METSO) that introduces pilot projects that use incentive based mechanisms relaying on the voluntarism of forest owners. The pilot projects test different types of compensating mechanisms. The compensation or payment for conservation contract is tied either only to the potential of forest revenue, or to that potential and to the conservation value (i.e. decayed wood, large aspen trees and such environmentally valuable elements have "a price tag"). In many cases, the sites that are most valuable for biodiversity conservation are not necessarily the most productive forestry areas (lots of tree species not used commercially, lots of decayed wood, long hauling distance, etc.). The new policy measures are hoped to bring about positive social and economic impacts through improvements in the acceptability of conservation among forest owners, and the cost effectiveness both for the state and for forest owners as a group. This study examined the factors that affect the acceptability of voluntary contracts of biodiversity conservation in NIPF and the amount of compensation required.

## 2 Material and Methods

#### 2.1 Choice Experiment Method

In order to analyse preferences for contract terms we applied the choice experiment method. The method involves respondents being presented with a number of choice sets consisting of two or more alternatives from which he/she is to choose the preferred alternative. Each alternative is described by various levels of a set of attributes, which are influenced by the chosen forest management strategy. Attributes can be quantitative or qualitative in nature, and the ability to combine these two types of data is one of the main benefits of the choice experiment approach.

Choice experiments are based on random utility theory and produce a wide range of information on trade-offs among the benefits provided by the choices (Adamowicz et al. 1997, 1998). The theory is based on probabilistic choice, where individuals are assumed to choose a single alternative, which maximises their utility from a set of available alternatives. Probabilistic choice models rely on random utility theory which describes the utility of each alternative (U) as the sum of systematic and error components. The systematic component, V, is a vector of individual and alternative specific attributes that are observable. The presence of an error component  $\varepsilon$  makes the choice random, and it includes all the impacts and factors affecting the choice that are not observable by the researcher (Louviere et al. 2000).

Random utility theory posits that an individual, n, chooses an alternative, i, from the choice set,  $C_n$ , if the indirect utility of i is greater than that of any other choice j. The following equation identifies this notion:

$$U_{in} > U_{jn} \Longrightarrow V_{in} + \varepsilon_{in} > V_{jin} + \varepsilon_{jn} \quad \forall \ j \neq i; i, j \in C_n (1)$$

Random utility theory describes the probability with which an alternative is chosen given its systematic and error components. The probability of individual n choosing an alternative i is the same as the probability that the utility of alternative iis greater than the utility of any other alternative of the choice set. Thus:

$$P(i) = P(V_{in} + \varepsilon_{in} > V_{jn} + \varepsilon_{jn}) \quad \forall \ j \neq i; i, j \in C_n$$
(2)

The conditional logit model is the most commonly used method in the analysis of multi-attribute choices. Assuming that the error components have an IID Gumbel distribution (Ben-Akiva and Lerman 1985, p. 104), the probability of choosing i is:

$$P(i) = \frac{\exp^{Vin}}{\sum_{j} \exp^{Vjn}}$$
(3)

The model is estimated using maximum likelihood estimation procedures and assumes a linearin-parameters functional form for the systematic portion of the conditional indirect utility function (Ben-Akiva and Lerman 1985).

Observing the choices made and the association of different attribute levels to monetary changes allows the estimation of changes in economic welfare. The compensating surplus (CS) for the case we examine can be written as:

$$CS = -\frac{V_{jn}^0 - V_{jn}^1}{\alpha}$$
(4)

where  $\alpha$  is the marginal utility of money, and  $V_{jn}^0$  and  $V_{jn}^1$  are the initial and new states of the resource. The initial state, or status quo, thus provides the basis for economic welfare analysis (Carson et al. 1994). Typically the marginal utility of money is derived from the parameter estimated in the choice model for some monetary attribute.

Choice experiment data was analysed using multinomial logit model with software programme Limdep8.0.

#### 2.2 Data Gathering

The data were collected using a postal survey to 3000 Finnish private forest owners in spring 2003. The sample was randomly collected from The Central Union of Agricultural Producers and Forest Owners' (MTK) and Forest Management Associations' register of landowners with more than five hectares of forest and who pay the full forestry levy. The sample included every 100th private forest owners across the whole country except Ahvenanmaa. Finland being bilingual, the questionnaire was translated into Swedish for those forest owners with an address in Swedish. After the first mailing, a reminder postcard and then a reminder with the questionnaire were sent. Some of the respondents were not found in the given address, were deceased or had sold their forest property, and after omitting these from the sample, the sample size was 2952. The response rate was 42%, which is abnormally low for this population. Therefore, researchers made a survey over a telephone to 100 randomly selected non-responding forest owners in November 2003. Most of the demographic characteristics were statistically similar to those of respondents. Non-respondents were more often farmers than respondents (30% and 22%, respectively). However, the higher percentage of farmers might be due to telephone survey: most of the calls were made in day-time, and if no-one answered the phone, the next person in the list was approached. Nearly 75 per cent named general disinterest in surveys or old age as a reason for non-respondence.

There were six series of survey questionnaires, each containing six choice sets. Each choice set included two contract alternatives for forest conservation that were described using five attributes, and a status quo alternative in which the level of conservation in private forests would not be increased. Respondents were instructed to choose their preferred alternative in each choice set.

The five attributes describing the alternative contract alternatives were: who initiated the conservation contract; the restrictions on forest use; the compensation per hectare annually; the duration of contract; and, the cancellation policy (Table 1).

#### Initiator

The first option was that the forest owner herself or himself is active in initiating the conservation contract. Conventionally, environmental organisations, the second option, have been active in initiating conservation actions, while the forest organisations, the third, have dealt with timber trading and extension of silvicultural practises. The new policy programme suggests the formation of a conservation trust that would be funded by voluntary payments for biodiversity conservation purposes, which was given as a fourth option of an initiator.

Attribute	Levels				
Initiator of the contract	Forest owner him/herself				
	Forest organisation				
	Environmental organisation				
	Conservation trust				
Restrictions on forest use	Small patches of forest protected				
	Nature management plan				
	No silvicultural practises allowed				
	Strict nature reserve				
Compensation/ha/year	0 euros				
	70 euros				
	140 euros				
	210 euros				
	280 euros				
	350 euros				
Duration of contract	5 years				
	10 years				
	30 years				
	100 years				
Cancellation policy	Forest owner can cancel				
	New owner can cancel				
	Binds also new owner				

Table 1. Attributes	used i	in the	study	and	their	levels.
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#### Restrictions

About a third of forest owners leave some small patches of forests unmanaged, so the small patches of forest protected would be an attractive option for many forest owners. The second option, a nature management plan would involve a voluntary plan that safeguards and enhances nature values in the forests but also allows harvesting. The third option for restrictions on forest use was a total ban on silvicultural practises. The most restrictive management option was a creation of a strict nature reserve that might impose restrictions on other uses as well as forestry.

#### Compensation per hectare per year

The amount of compensation proposed varied between 0 to 350 euros. Using the word payment instead of compensation might be more appropriate when the forest owner is paid for biodiversity services rather than compensated for lost timber revenue. However, in the following welfare analysis, we are calculating how much compensation would be required to keep forest owner's welfare intact.

#### Duration of a contract

The levels of the duration of contract ranged from five

years to one hundred years that would already cover in average three generations of forest ownership.

#### Cancellation policy

The levels of cancellation policy varied in terms of who is allowed to cancel the contract. One level was that the forest owner who enters into a contract might cancel it and, naturally, return the compensation due. Alternatively, the contract would bind the forest owner but a new owner would be allowed to cancel the contract. Lastly, the contract would also bind the new forest owner.

# 2.3 Heterogeneity of Preferences for Nature Conservation

Many earlier empirical valuation studies ignore the taste variations across the respondents in the sample. If the quantity of environmental good demanded varies significantly between individuals this preference heterogeneity might lead to misinterpretation of results. A growing number of recent studies have accounted explicitly for heterogeneity in preferences through the use of econometric techniques. For example, Siikamäki (2001) used random coefficient models to demonstrate preference heterogeneity for conservation policy among the population. Adamowicz and others (1997) interacted respondent specific socio-demographic characteristics with the design attributes to incorporate the impact of different preferences on choices.

Assuming that the supply of forest stands offered for conservation contracts exceeds the demand, a conservation policy where forest owners take an initiative themselves might prove to be a cost effective conservation policy for the country. Those forest owners who internalise some of the nature values into their own objective function for forest ownership might enter into a conservation contract at a lower price per hectare than those with no conservation interests. Alternatively, some forest owners might not take advantage of the potential harvesting revenue in any case, and thereby a conservation contract might offer them a way of profiting from forest ownership without timber sales. Finnish forest owners have been found to have varying goals in their forest ownership (Karppinen et al. 2002), and this heterogeneity might bring about both social and economic benefits in forest conservation.

In this study, the heterogeneity in forest owners' attitudes towards forest conservation and incentive based policy mechanisms were taken into account by examining the choices for the status quo alternative. The status quo provides respondents with something familiar and also provides a means to say that no change is preferable. When a respondent always chooses the status quo alternative he or she indicates a preference for the status quo over all other alternatives that have been available. In this study, a consistent choice of the status quo alternative indicates that the respondent is not willing to enter into any conservation contract, at least within the contract term levels presented in the choice sets. On the other hand, the respondents who have chosen either of the contract alternatives at least once indicate willingness to negotiate a conservation contract if the terms are agreeable. In this study, the heterogeneity of respondents was examined by modelling two separate models: one for all respondents and one excluding status quo respondents.

# **3** Results

## 3.1 Sample Characteristics

The sample characteristics represent well those of the largest Finnish forest owner study by Karppinen and others (2002). The average size of forest property was 42 hectares. The respondents were on average 58 years old, 41 percent had retired while 28 percent were employees and 22 percent were farmers. Four fifths of respondents were male. Differences between the respondents in terms of choosing the status quo were tested. Those who always chose the status quo are older than the respondents in average, and thereby also their educational level is lower and a larger proportion is retired. Gender or the characteristics of forest property were not statistically significant. One third of them were of the opinion that the area of forest conservation should be lower while the overall percentage was 22 percent. Always choosing the status quo thus seems to reflect genuine preferences rather than be an easy way out from answering.

### 3.2 Estimation of Multinomial Logit Models

About one third of the respondents always selected the status quo alternative in their choices. Two models were thus estimated; one using all the observations in the data set collected, and one from where the respondents who had always chosen the status quo had been omitted. The statistical results of the separate multinomial logit models are shown in Table 2. The co-efficients mark the effect of that parameter on the probability of an alternative to be chosen.

The status quo alternative was assigned the alternative specific constant (ASC). The positive and statistically significant ASC indicates strong preferences for no additional conservation. The compensation parameter estimate is also positive for both models, indicating that the higher the compensation in the alternative, the higher the probability of it being chosen.

The different compensation demands might also be due to different opportunity costs in forestry in different parts of the country as the potential for forest growth varies according to vegetation

Variable	Parameters (SE)				
	All data	Status quo respondents excluded			
Alternative specific constant for the status quo Compensation	(	762)0.7756***(0.0967)003)0.0047***(0.0004)			
InitiatorForest ownerInitiatorForest organisationInitiatorEnvironmental organisationInitiatorConservation trust	0.0573 (0.0	607)         0.6294***         (0.0722)           664)         -0.0752         (0.0791)           642)         -0.2890***         (0.0796)           -0.2650         -			
Restriction on useSmall patches conservedRestriction on useNature management planRestriction on useNo silvicultureRestriction on useStrict nature reserve	0.2373*** (0.0	580)         0.5450***         (0.0729)           695)         0.3693**         (0.0830)           660)         -0.2440**         (0.0787)           -0.6706         -			
Duration of contract5 yearsDuration of contract10 yearsDuration of contract30 yearsDuration of contract100 years	0.2865*** (0.0	592)         0.6432***         (0.0770)           6609)         0.3328***         (0.0721)           637)         0.0474         (0.0756)           -1.0234         -			
Cancellation policyPresent owner can cancelCancellation policyNew owner can cancelCancellation policyBinds also new owner	(	497)         0.3080***         (0.0594)           537)         0.1319**         (0.0610)           -0.4400         -			
Log-likelihood $\rho^2$	-2490.18 0.1889	-1555.72			

**Table 2.** Estimated parameters of multinomial logit models (and standard errors) using all data and the data with status quo respondents excluded.

\*\*\* significant at p < 0.01; \*\*significant at p < 0.05, \* significant at p < 0.10

zones. However, the two data had no statistical difference in terms of respondents' residential region in chi-square test. Also, interacting a respondentspecific variable of the three regions in Finland with the compensation variable turned out to be statistically insignificant. However, opportunity costs per each property specifically are not available in the data.

All the other variables were effects coded. Effects coding provides a useful way of including qualitative attributes into the analysis. The main effect of an attribute with L levels can be defined by creating L-1 effects coded variables which take values of 1, 0 or -1 depending if the alternative contains the value of the new variable (see e.g. Louviere et al. 2000). At least one of the levels was statistically significant for all the variables. Nearly all the signs of variable levels were the same in both models, and the preference order of the levels is likewise similar. "Forest owner" was the most preferred alternative for the initiator of the contract, while the "environmental organisations" were least favoured. Respondents were willing to conserve small patches of forest or manage their forest according to a nature management plan, rather than take more restrictive measures. Short contract periods were preferred to longer ones, with the practically permanent one, hundred years, being a highly unpopular choice. Respondents also preferred flexibility in decision, opting to have a possibility of withdrawing from the contract at their will.

What differs between the two models is the magnitude of the ASC co-efficient. Those who always chose the status quo having been excluded, the status quo ASC is naturally much smaller for the respondents who are willing to make a contract. They also have a higher compensation demand, and the role of themselves as the initiators of the contract is more important. Also the sovereignty in decision making shows in the high preference for a flexible cancellation policy.

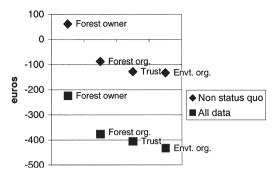
### 3.3 Welfare Analysis

The welfare analysis shows that the terms of the contract are of great importance to forest owners as the demand for compensation rises manyfold with undesirable factors. The base scenario was selected to have the forest owner as the initiator of the contract, the contract binds a new as well as the present forest owners, small patches are protected and the duration of contract is 10 years. In this base scenario the impact on forest owners welfare is -224 euros per hectare annually when using all data. So at least this amount should be paid to a forest owner on average as compensation for biodiversity conservation services to hold his or her welfare constant. However, the welfare impact for the same contract but estimated using the "Status quo respondents excluded" model results in a positive figure of +62 euros per hectare annually. Thus these forest owners would have a positive welfare impact of the described contract. In other words, these forest owners would be made better off with the introduction of the contract to the extent of 62 euros per hectare per annum.

If the contract terms are changed, the welfare impact shifts accordingly. For example, if any other agent rather than the forest owner initiates the contract, the welfare change declines drastically for both data, being now negative also for the non status quo group (Fig. 2). Similarly, if the restrictions on forest use are changes from only small patches to be protected to larger areas left outside silvicultural management, the welfare impact would be -400 euros and -105 euros per hectare annually, respectively.

## 4 Discussion

This study examined Finnish NIPF owners' preferences for biodiversity conservation in private forests. The choice experiment method was applied to study the trade-offs between compensation and the terms of voluntary conservation contracts. It is the first application of choice experiment method in valueing contract terms of environmental policy. Heterogeneity among forest owners was examined by modelling sepa-



**Fig. 2.** Changes in welfare by data when the initiator of contract is changed while the other variables are held at the base scenario level.

rately those respondents who had made choices between the status quo situations and the contract alternatives. The results contribute to the improved design of voluntary mechanisms in biodiversity conservation policy.

The response rate was abnormally low for the population of private forest owners. A phone survey to 100 non-respondents revealed that in most demographic characteristics non-respondents were similar to the respondents, but the non-respondents were more often farmers. As the non-respondents preferred for more biodiversity conservation in private forests, but on the other hand had themselves set aside less of their own property, it is inconclusive whether they would enter into conservation contracts and at what price. The focus in conservation contracts is on the initiative and interest of forest owners, and the passive forest owners would not affect market as such.

The results show how the welfare of forest owners shifts when the contract terms are changed. To keep the forest owners' welfare constant before and after the conservation contract, the owner would have to be compensated the amount of the welfare change. Welfare does not only reflect the economic potential of the forest, but also the other values, e.g. for biodiversity conservation or for forest owners' sovereignty. In a base scenario the forest owner was assumed to be the initiator of the contract that would require only small patches of forest to be protected, and would also bind new forest owners over its duration of ten years. For all respondents, the average demand for compensation would be around 224 euros. This is higher than the average annual revenue from timber sales from a hectare over the rotation period. However, if the respondents who were not willing to choose a conservation contract are excluded from the data, the average welfare impact would be +62 euros. This is not surprising in that many forest owners are already leaving some patches of their forest intact without getting any compensation of it (Horne et al. 2004). However, if the restrictions on forest use were more severe involving a larger area to be left outside silvicultural management, the welfare impact on this group would be -105 euros.

In the Finnish pilot projects testing different compensation mechanisms, the compensation or payment for conservation contract is at least partly tied to the potential of forest revenue. The age structure, species composition and annual growth would have an impact on the value of the forest stand, and thus on the opportunity cost of setting it aside from forestry for conservation. Data on forest stands of the properties were not available. However, inspection of the three regions, Southern, Western and Northern Finland, revealed no statistical difference in compensation demand. Thus the opportunity cost does not seem to hold the balance of power on whether a land owner would enter into a conservation contract or not.

For a conservation policy to be socially accepted and cost effective for both the government and the forest owners as a group, the heterogeneity in forest owners' preferences and goals for the forest holding should be taken into account. An earlier study by Jokinen and others (1997) showed that while forest owners might indicate positive attitudes towards nature conservation in general, concrete action on the part of conservation authorities would meet opposition. Voluntary mechanisms might convey positive attitudes into action by securing the power of decision with forest owners. Our results stress the importance of land ownership to many forest owners, and heterogeneous attitudes to biodiversity conservation. Instead of using a top down approach of imposing a conservation status on a NIPF holding, those forest owners willing to protect parts of their forest with biodiversity values could be

allowed to enter voluntarily into contracts with environmental officials of the state. This might result in a more cost-effective option.

The goal of nature conservation is to secure the protection of nature values in situ. Strict nature reserves provide a secure core for conservation networks and present a low risk level in the stability of conservation status. Considering only ecological values, the acquisition of forestland by the state for strict biodiversity conservation purposes would seem like an attractive option. However, the optimal choice of conservation policy and implementation mechanism is a complex matter of trade-offs between ecological values and socioeconomic considerations. Tailoring the policy mechanisms to suit the ecological requirements in a cost-effective and socially acceptable manner is a challenging task for the policy makers.

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